# Plant Transformation Pathways of Energetic Materials (RDX, TNT, DNTs)

Jong Moon Yoon, David J. Oliver and Jacqueline V. Shanks *Iowa State University Ames, IA* 

2,4,6-trinitrotoluene (TNT) and hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX) are commonly found in surface soil and groundwater at ammunition-production and military-training sites. Approximately 2,000 US Department of Defense facilities are contaminated with explosives both in soil and in groundwater (Medina *et al.*, 2003). Dinitrotoluenes (DNTs), such as 2,4-dinitrotoluene (2,4-DNT) and 2,6-dinitrotoluene (2,6-DNT), contaminate the sites as by-products of TNT. TNT and DNTs are classified as nitroaromatic explosives having aromatic ring structures, whereas RDX is a nitramine explosive possessing N-nitro groups (Hannink *et al.*, 2002). The explosives are transported to soil and groundwater after open detonation, seepage and/or improper disposal at military and munition-production sites. TNT and DNTs have higher octanol-water partition coefficient (Kow) values than does RDX, suggesting that they are strongly bound to soil organic matter, whereas RDX is mobile as a result of poor sorption. The explosives are not volatile due to their low vapor pressures. Physical and chemical properties of the explosives are shown in Table 1.

Phytoremediation is a promising technology using plants to clean up contaminated soil and groundwater in situ because of low cost of maintenance and operation, and public acceptance.

Several studies have reported abiotic methods of treatment of explosives, such as incineration, carbon adsorption, alkaline hydrolysis, and catalytic and advanced oxidation (Garg *et al.*, 1991; Rodgers and Bunce, 2001). Harmful by-products, requiring further treatment, and transport of contaminated soils or groundwater have drawn attention to bioremediation. Phytoremediation is a promising technology using plants to clean up contaminated soil and groundwater *in situ* because of low cost of maintenance and operation, and public acceptance (Schnoor *et al.*, 1995). The fact that plants are able

Table 1. Chemical and physical properties of TNT, RDX, and DNTs.				
	TNT	RDX	2,4-DNT	2,6-DNT
Molecular weight	227.15	222.26	182.14	182.14
Molecular formula	$C_7H_5N_3O_6$	$C_3H_6N_6O_6$	$C_7H_6N_2O_4$	$C_7H_6N_2O_4$
$Log\; K_{ow}$	1.6-1.84	0.81-0.87	1.98	1.9-2.10
Solubility in water (mg/L)	100	42	270-273	910
Vapor pressure (mm Hg)	1.99×10 <sup>-4</sup>	1.0-4.0×10 <sup>-9</sup>	$1.47 \times 10^{-4}$	5.67×10 <sup>-4</sup>
Henry's constant (atm-m <sup>3</sup> /mole)	4.57×10 <sup>-7</sup>	1.2×10 <sup>-5</sup>	1.3×10 <sup>-7</sup>	9.26×10 <sup>-8</sup>
O <sub>2</sub> Molecular structure	NO <sub>2</sub>	O <sub>2</sub> N NO <sub>2</sub>	CH <sub>3</sub> NO <sub>2</sub>	O <sub>2</sub> N NO <sub>2</sub>

Data from Yinon and Zitrin (1993), Talmage et al. (1999) and HSDB (2000).

to accumulate metals to high concentrations in their tissues is well known (Salt et al., 1998). Phytoremediation research has been conducted on organic pollutants ranging from pesticides, e.g. atrazine (Burken and Schnoor, 1997), to industrial pollutants such as trichloroethylene and polycyclic aromatic hydrocarbons (PAH) (Newman et al., 1997; Paquin et al., 2002; Vervaeke et al., 2003). The transformation products of xenobiotics by plants are less toxic than parent compounds. In addition, root exudates are reported to enhance microbial activity for degradation of xenobiotics (Miya and Firestone, 2001; Anderson et al., 1993). The cost for soil remediation is \$10 to \$100 per cubic meter whereas vegetative cleanup of contaminated soils costs only \$0.02 to \$1 per cubic meter (Cunningham et al. 1995).

This short review summarizes the toxicity of energetic materials (TNT, RDX, and DNTs) and pathways of transformation by plants.

## Toxicity of the Explosives

TNT and its degradation products have been reported to be mutagenic and toxic to several organisms. Survival of the midge (Chironomus tentans) decreased significantly after exposure to 200 mg/kg of TNT, 1,3,5,-trinitrobenzene (TNB), and 2,4-diamino-6-nitrotoluene (2,4-DANT); and the amphipod (*Hyalella azteca*) was more susceptible to TNT, TNB, and 2,4-DANT than the midge (Steevens et al., 2002). Gogal et al. (2002) reported that northern bobwhite quail showed decreases in total red blood-cell counts and plasma protein as dietary TNT intake increased, and they determined a low observedadverse-effect level of 178 mg of TNT per kg of weight per day. Survival and growth of two freshwater invertebrates were not affected after a 10-day exposure to 1,000 mg of RDX per kg of sediment (Steevens *et al.*, 2002). The growth and survival of benthic invertebrates, *Neanthes arenaceodentata* and *Leptocheirus plumulosus*, were not affected by exposure up to 1,000 µg RDX per kg dry weight of sediment (Lotufo *et al.*, 2001). Inhibition to growth and reproduction of adult earthworms can occur at less than 95 mg of RDX per kg of artificial soil (Robidoux *et al.*, 2000, 2001), but acute toxicity was not observed up to 756 mg per kg dry soil for RDX.

Gong et al. (2002) investigated the influence of RDX on indigenous microbial activities. They measured soil dehydrogenase activity, potential nitrification activity, heterotrophic nitrogen fixation activity, substrate-induced respiration, and basal respiration for 12 weeks. Significant reductions (up to 30% of control) in these parameters were observed in RDX-spiked soil. In the case of a luminescent marine bacterium (Vibrio fischeri), the EC<sub>50</sub> value of RDX (116 mg/L) was above the solubility in water (42 mg/L for RDX) after incubation periods of 90 min (Drzyzga et al., 1995).

Neither 2,4-DNT nor 2,6-DNT were mutagenic with the Ames assay, whereas their hydroxylamine isomers proved to be mutagenic (Padda *et al.*, 2003). Using the uptake response of H4IIE rat hepatoma cell cultures to neutral red, the NR<sub>50</sub> values were 45 mg/L for 2,4-DNT, 50 mg/L for 2,6-DNT, and 7 mg/L for TNT, suggesting dinitrotoluenes are less cytotoxic than TNT (Mitchell and Burrows, 1995).

TNT was toxic to hybrid poplars at a concentration of 5 mg/L in hydroponic solution (Thompson *et al.*, 1998) and at 50 mg/kg soil there were adverse effects on germination and seedling growth of cress and turnip (Gong *et al.*, 1999). Alfalfa did not grow at 0.55 mM (100 mg/kg) 2,4-DNT in soil (Dutta *et al.*, 2003), and lettuce was more sensitive than wheat, mustard, and lentil, indicating that phytotoxic effects of nitroaromatic explosives depend on plant species (Picka and Friedl, 2004). The highest non-observed adverse effect concentrations (NOAEC) for the growth of lettuce were 20 mg/kg for TNT, 2 mg/kg for 2,4-DNT and 10 mg/kg for 2,6-DNT. Hydroponic toxicity of RDX to maize and wheat was estimated to be 21 mg/L RDX, while soybean and sorghum did not show a toxic effect up to 21 mg/L for 30-day exposures (Chen, 1993). RDX was not toxic to hybrid poplars up to 21 mg/L (Thompson, 1997).

## MECHANISMS OF DEGRADATION OF XENOBIOTICS BY PLANTS

Prior to the introduction of xenobiotics to plant cells, they must be taken up through the roots. Several studies reviewed predictive relationships between the uptake rate of a compound and its physical-chemical properties (Briggs  $\it et al.$ , 1982; Burken and Schnoor, 1998). Root uptake and translocation of the compounds are related to the logarithm of the octanol-water partition coefficient, log  $K_{ow}$ . Root concentration factor (RCF), defined as the the concentration sorbed to the roots divided by the concentration in the aqueous phase, is generally proportional to the log  $K_{ow}$  value. The relationship is proposed as follows:

 $log (RCF-3.0) = 0.65 log K_{ow} - 1.57 by Briggs et al. (1982)$  $log (RCF-0.82) = 0.77 log K_{ow} - 1.52 by Burken and Schnoor (1998)$  The transpiration stream concentration factor (TSCF) is calculated as the concentration in the transpiration stream divided by the aqueous concentration. The values of TSCF for various chemicals show Gaussian distribution curves over the range of log  $K_{ow}$  values, indicating that hydrophilic compounds (log  $K_{ow}$  < 1.8) are not able to pass through lipid membranes of roots, whereas hydrophobic compounds (log  $K_{ow}$  > 3.8) tend to bind strongly to root tissues and are not then translocated to shoots (Dietz and Schnoor, 2001). The relationship between TSCFs and log  $K_{ow}$  is proposed as follows:

```
TSCF = 0.784 exp{-(\log K_{ow} - 1.78)^2/2.44} by Briggs et al. (1982)
TSCF = 0.756 exp{-(\log K_{ow} - 2.50)^2/2.58} by Burken and Schnoor (1998)
```

Enzymatic transformation of xenobiotics by plants follows the green-liver model and involves three steps. First, the foreign compounds taken up by plants are transformed by enzymes such as cytochrome P450, carboxylesterases, and peroxidase (Sandermann, 1994). Secondly, the transformed xenobiotics are conjugated with D-glucose, glutathione, or amino acids (Komoba *et al.*, 1995) by enzymes such as glutathione *S*-transferases, glucosyltransferases and malonyltransferases, resulting in either soluble or insoluble products. The third step is storage and compartmentation; the soluble compounds are stored in vacuoles or as cell-wall materials by further processing, and the insoluble compounds are generally assumed to be stored in the cell wall (Schroder and Collins, 2002).

## UPTAKE OF THE ENERGETIC MATERIALS BY PLANTS

Nitroaromatic explosives showed different uptake and fate in plant systems than nitramine explosives. According to Thompson *et al.* (1998), 95% of the TNT was removed from solution in less than 24 h by hybrid poplar, whereas 71% of the RDX was removed from hydroponic solution in 7 days (Thompson *et al.*, 1999). The uptake of both RDX and TNT from soil was slower than in the hydroponic systems because of decreased bioavailability in soil. Bush beans took up less than 16% of RDX in soil after 60 days; in contrast 60% was removed from solution after 7 days (Harvey *et al.*, 1991).

Over 60% of radioactivity of <sup>14</sup>C-RDX taken up by hybrid poplars was found in the leaves after 2 days. In contrast, 78% of radioactivity of <sup>14</sup>C-TNT taken up by the poplars remained in the roots after the same exposure time (Thompson *et al.*, 1998), suggesting that RDX is translocated more readily. In addition, an overall low recovery of RDX with no significant mineralization by plants suggested that the final transformation products are volatile compounds (Just and Schnoor, 2000). Recently, poplar nodule cultures were reported to mineralize RDX under sterile conditions (Van Aken *et al.*, 2004).

Regarding DNTs, knowledge of uptake by plants and transformation products is limited compared to information on TNT and RDX. Best *et al.* (2001) applied wetland systems to remove explosives from groundwater at ammunition plants, resulting in average removals of 58% and 61% for 2,4-DNT and 2,6-DNT, respectively, in a 115-day operation at the Volunteer Army Ammunition Plant, Chattanooga, TN. Todd and Lange (1996) observed that 67% of 2,4-DNT from soil was removed in a phytoremediation system using parrot feather (*Myriophyllum brasiliense*). They found 4-amino-2-nitrotoluene (4A2NT) in the

plant tissues after 90 h of treatment prior to 2-amino-4-nitrotoluene (2A4NT) which was detected after 190 h of exposure. However, other transformation products of the DNTs, as well as their fate in plants, are unknown.

## Transformation Pathways

## 2,4,6-trinitrotoluene (TNT)

Subramanian and Shanks (2003) proposed the TNT transformation pathway by plants based on experiments with periwinkle (*Catharanthus roseus*) and parrot feather, as shown in Figure 1.

Two monoamino compounds [2-amino-4,6-dinitrotoluene (2ADNT) and 4-amino-2,6-dinitrotoluene(4ADNT)] have been found as the primary reduction products by plants (Palazzo and Leggett 1986; Thompson *et al.* 1998; Bhadra *et al.*, 1999a). Diaminotoluenes (2,4-diamino-6-nitrotoluene and 4,6-diamino-2-nitrotoluene) and azoxy compounds were observed under strong reducing conditions and by the condensation of hydroxylamines, respectively (Pavlostathis *et al.*, 1998; Sens *et al.*, 1998; Thompson *et al.*, 1998).

As for oxidative transformation of TNT in plant systems, Bhadra *et al.* (1999b) isolated six oxidized metabolites such as 2-amino-4,6-dinitorbenzoic acid, 2,4-dinitro-6-hydroxy-benzyl alcohol, 2-*N*-acetoxyamino-4,6-dinitrobenzaldehyde, 2,4-dinitro-6-hydroxytoluene, and two binuclear metabolites from azoxytetranitro toluenes. In addition, they showed that oxidation of TNT by the plant could occur before the reductive transformation. This was based on results where monoamino compounds were added to plants and the oxidized metabolites of TNT were not produced. To date, oxidized metabolites have only been found in parrot feather; they were not detected in *Catharanthus* or *Arabidopsis* (Subramanian, 2004).

2-hydroxylamino-4,6-dinitrotoluene (2HADNT) and 4- hydroxylamino-2,6-dinitrotoluene (4HADNT) were observed following reduction of nitro groups of TNT in non-axenic and aquatic plant systems (Pavlostathis *et al.*, 1998; Wang *et al.*, 2003). Measurement of hydroxylamines was difficult due to their instability. Wang and Hughes (1998) developed an efficient assay for hydroxylamines by derivatization with acetic anhydride. Recently, these hydroxylamines were shown to be present in axenic hairy roots of *Catharanthus* and axenic *Arabidopsis* seedlings (Subramanian, 2004; Subramanian *et al.*, 2005). The hydroxylamines are considered the first transformation products resulting in other metabolites of TNT by reduction, oxidation, conjugation, and polymerization (Subramanian and Shanks, 2003; Wang *et al.*, 2003).

The transformed products of TNT are further conjugated and sequestered in plant cells. Over 80% of the TNT label was associated with plant biomass, suggesting that the labeled carbon from TNT was sequestered in the plant tissues (Harvey *et al.*, 1991). Thompson *et al.* (1998) showed that 75% of the radioactivity of <sup>14</sup>C-TNT was present in unextractable and bound residues in the poplar roots. Bhadra *et al.* (1999a) characterized the four conjugates of TNT metabolites with a 6-carbon moiety by *Catharanthus roseus* and *Myriophyllum aquaticum*. They found that two of them have molecular structures similar

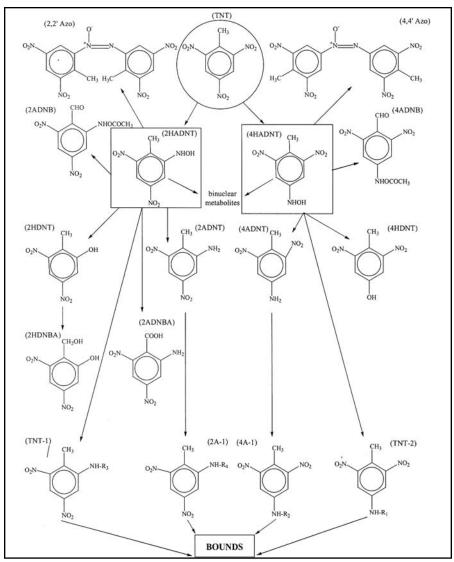


Figure 1. TNT transformation pathway by Subramanian and Shanks (2003). Abbreviations—TNT: 2,4,6-trinitrotoluene; 2ADNT: 2-amino-4,6-dinitrotoluene; 4ADNT: 4-amino-2,6-dinitrotoluene; 2HADNT:2-hydroxylamino-4,6-dinitrotoluene; 4HADNT: 4-hydroxylamino-2,6-dinitrotoluene; 4,4'Azo: 2,2',6,6'-tetranitro-4,4'-azoxytoluene; 2,2'Azo: 4,4',6,6'-tetranitro-2,2'-azoxytoluene; 2HDNT: 2-hydroxy-4,6-dinitrotoluene; 4HDNT: 4-hydroxy-2,6-dinitrotoluene; 2ADNB: 2-N-actamido-4,6-dinitrobenzaldehyde; 2ADNB: 4-N-actamido-2,6-dinitrobenzaldehyde; 2HDNBA: 2-hydroxy-4,6-dinitrobenzyl alcohol; and 2ADNBA: 2-amino-4,6-dinitrobenzoic acid. TNT-1, TNT-2, 2A-1 and 4A-1 represent conjugates with six carbon sugars (R1, R2, R3, and R4).

to that of 2ADNT (labeled TNT-1 and 2A-1) and the others were similar to 4ADNT (TNT-2 and 4A-2), indicating that the monoamines were precursors to the conjugates. Recent studies have elucidated these TNT conjugates. The conjugates of TNT metabolites by tobacco cell cultures are formed by conjugation of glucose on the hydroxylamine group of either 2HADNT or 4HADNT, and various diglycoside conjugates with gentiobioside or sophoroside forms were identified, including monoglycosides (Vila *et al.*, 2005). In precursor-feeding studies, Subramanian (2004) and Subramanian *et al.* (2005) found evidence for conjugation of monoamines and hydroxylamines with plant sugars.

## Hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX)

Studies on the transformation of RDX by plants are rare, whereas several microbial transformation pathways have been proposed and some established. After being taken up and translocated to leaf tissues, direct photolysis of RDX in the leaves is a feasible fate under natural sunlight. Just and Schnoor (2004) proposed the photodegradation pathway of RDX by reed canary grass, as shown in Figure 2. They identified ring-cleavage products, such as nitrous oxide (N2O) and 4-nitro-2,4-diazabutanal in leaves under simulated sunlight, including nitrite (NO<sub>2</sub>-) and formaldehyde (CH<sub>2</sub>O) in solution. Van Aken et al. (2004) proposed three processes for the pathway of degradation of RDX by using poplar tissue cultures and crude extracts from leaves, as shown in Figure 2. First, reduction products such as hexahydro-1-nitroso-1,3-dinitro-1,3,5-triazine (MNX) and hexahydro-1,3-dinitroso-5-nitro-1,3,5-triazine (DNX) were produced by intact plant cells regardless of light. In the second step, the reduced metabolites were further transformed to formaldehyde and methanol, both in crude extracts and in intact cultures under light. In the final step, light-independent mineralization of one-carbon metabolites by intact plant cultures, but not crude extracts, occurred. Some transformed products may be re-incorporated into plant cells. Formaldehyde may be conjugated by plant enzymes to form compounds like S-formyl-glutathione (Just and Schnoor, 2004). Small quantities of CO<sub>2</sub> produced by degradation of RDX by plants may be re-assimilated by photosynthesis (Van Aken et al., 2004).

## Dinitrotoluenes (DNTs)

In contrast with bacterial systems, little information is available on the transformation of DNTs by plants. We are aware of only one study: monoamino isomers, 2A4NT and 4A2NT, were reductive transformation products in plants (Todd and Lange, 1996). The bacterial reduction of dinitrotoluenes can take place under aerobic and anaerobic conditions, resulting in the production of monoamines isomers (Hughes *et al.*, 1999). Hydroxylaminotoluenes and dihydroxylaminotoluenes were produced anaerobically in cell cultures of *Clostridium acetobutylicum* (Hughes *et al.*, 1999). Further transformed products, aminohydroxylaminotoluenes and diaminotoluenes, were observed in the cell extracts. *Hydrogenophaga palleronii* and *Burkholderia cepacia* produced oxidative intermediates and mineralized DNTs to CO<sub>2</sub> by mono- or dioxygenases (Nishino *et al.*, 1999). The bacteria converted 2,6-DNT into 3-methyl-4-nitrocatechol with release of nitrite, and then 2-hydroxy-5-nitro-6-oxohepta-2,4,-dienoic acid and 2-hydroxy-5-nitropenta-2,4,-dienoic acid (Nishino *et al.*, 2000).

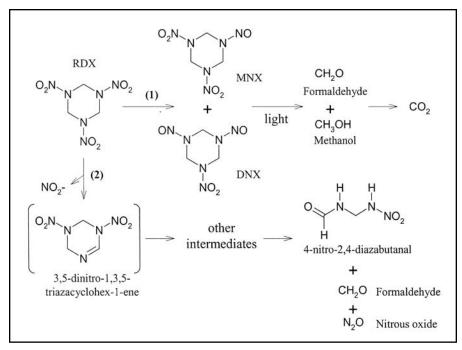


Figure 2. RDX degradation pathways proposed by Van Aken *et al.* (2004) and Just and Schnoor (2004). The bracketed compound was not observed. Abbreviations—RDX: hexahydro-1,3,5-trinitro-1,3,5-triazine; MNX: hexahydro-1-nitroso-1,3-dinitro-1,3,5-triazine; and DNX: hexahydro-1,3-dinitroso-5-nitro-1,3,5-triazine.

Transgenic tobacco plants expressing nitroreductases of Enterobacter cloacae showed enhanced ability to tolerate and remove TNT at high concentration.

## Transgenic Plants and Gene Expression

In the past 6 years there has been significant activity in using genetic approaches to enhance transformation and to reduce phytotoxicity of energetic materials. Genetically modified plants expressing bacterial genes have been developed for phytoremediation. Transgenic tobacco plants expressing nitroreductases of *Enterobacter cloacae* showed enhanced ability to tolerate and remove TNT at high concentration (0.25 m*M*), which is toxic to wild-type tobacco (Hannink *et al.*, 2001). Another transgenic tobacco line expressing pentadrythritol tetranitrate reductase from the bacterium showed better germination and growth in the presence of TNT (0.05 m*M*) than did wild-type plants (Rosser *et al.*, 2001; French

et al., 1999). In addition, these researchers showed enhanced RDX removal in tobacco engineered with an XplA cytochrome P450 from *Rhodococcus rhodochrous* (unpublished results). Clearly, genetic modification with microbial redox enzymes has the potential to enable faster transformation of TNT and RDX and reduced phytotoxicity.

Transcriptomic studies are providing clues to endogenous plant genes involved in transformation. Specific genes such as those for glutathione-S-transferases and cytochrome P450 in Arabidopsis were proposed by Ekman et al. (2003) to be involved in transformation of explosives. They used serial analysis of gene expression (SAGE) to compare 14-day-old Arabidopsis, exposed to 15 mg/L of TNT after 24 h, to untreated plants. A glutathione-S-transferase was found to be induced up to 27-fold. Among the highly induced genes were those encoding cytochrome P450 (CYP81D11-A-TYPE), an ABC transporter that is known to expend ATP energy to transport hydrophobic molecules into or out of the cytoplasm, and a 12-oxophytodienoate reductase having high homology to nitroreductases of Enterobacter sp. (Ekman et al., 2003). However, as noted previously, oxidative compounds were not found in Arabidopsis (Subramanian, 2004), thus the role of P450s in transformation pathways in Arabidopsis is unclear. In microarray experiments, Arabidopsis gene expression was monitored after long-term exposure (10 days) to various concentrations of TNT (Mentewab et al., 2005). In response to TNT amendment, fiftytwo genes were upregulated and forty-seven were downregulated, many of which have cell-defense and detoxification functions. Glutathione-S-transferases and cytochrome P450s were not found to be significantly upregulated in this study. Most of the genes differentially expressed were observed at the higher concentration of TNT amendment  $(10 \,\mu M)$  and genes expressed at 1 and 10  $\mu M$  rarely overlapped. They confirmed the gene expressions of pathogenesis-related protein-1 precursor, DNA-binding protein, and ABC transporter-like protein by real-time PCR analysis.

The transcriptome studies provide clues to genes that may be involved in TNT transformation. Upregulation of some of the genes may be the result of a generalized stress response without synthesis of enzymes involved in the TNT phytotransformation pathway or in a reduced phytotoxicity response. Reverse-genetics approaches using the genes identified should enable further clarification of the transcriptome results. In a forward-genetics approach, ten activation-tagged *Arabidopsis* mutant lines showing significantly better germination rates than the wild type on the TNT-amended medium were isolated from 300,000 mutant seeds (Moon *et al.*, 2004).

Selection of high-performing native plants, engineering plants with enhanced transformation capabilities, identifying the fate of transformation products in plants, and designing the external variables to operate a more effective phytoremediation process are all dependent on a knowledge base of the genetic structure, enzymatic structure, and biochemical reaction pathways. The genetic approaches discussed here will enable the design of effective strategies for remediation of energetic materials in the future.

## Conclusions and Future Directions

Plants can remove contaminants from soil and groundwater, and transform them into less harmful compounds. Based on information on transformation pathways and gene

## Research on the post-harvest fate of explosives is required.

expression, further studies on metabolic engineering and genetic modifications may make plants tolerant to higher concentrations of xenobiotics by inducing faster rates of uptake and using less toxic metabolic pathways. In addition, the explosives taken up by plants can be released by action of water—e.g. rain and river—and thus may be returned to the environment as hazardous contaminants. Research on further treatments and the post-harvest fate of explosives is required. Information about phytoremediation of dinitrotoluenes is lacking compared to that for TNT and RDX; thus, it also merits further investigation.

#### Acknowledgment

This research was supported in part by the US Department of Defense through the Strategic Environmental Research and Development Program (SERDP), Project CU-1319.

#### References

- Anderson TA et al. (1993) Bioremediation in the rhizosphere. Environmental Science and Technology 27 2630-2636.
- Best EPH et al. (2001) Tolerance towards explosives and explosives removal from groundwater in treatment wetland mesocosms. Water Science and Technology 44 515-521.
- Bhadra R et al. (1999a) Confirmation of conjugation processes during TNT metabolism by axenic plant roots. Environmental Science and Technology 33 446-452.
- Bhadra R et al. (1999b) Characterization of oxidation products of TNT metabolism in aquatic phytoremediation systems of Myriophyllum aquaticum. Environmental Science and Technology 33 3354-3361.
- Briggs GG et al. (1982) Relationships between lipophilicity and root uptake and translocation of non-ionised chemicals by barley. Pesticide Science 13 495-504.
- Burken JG SchnoorJL (1997) Uptake and metabolism of atrazine by poplar trees. Environmental Science and Technology 31 1399-1406.
- Burken JG Schnoor JL (1998) Predictive relationships for uptake of organic contaminants by hybrid poplar trees. Environmental Science and Technology 32 3379–3385.
- Chen D (1993) Plant Uptake and Soil Adsorption of RDX. Master's Thesis: University of Illinois.
- Cunningham SD et al. (1995) Phytoremediation of contaminated soils. Trends in Biotechnology 13 393-397.
- Dietz AC Schnoor JL (2001) Advances in phytoremediation. Environmental Health Perspectives 109 163-168.
- Drzyzga O et al. (1995) Toxicity of explosives and related compounds to the luminescent bacterium vibrio fischeri NRRL-B-11177. Archives of Environmental Contamination and Toxicology 28 229-235.

- Dutta SK *et al.* (2003) Enhanced bioremediation of soil containing 2,4-dinitrotoluene by a genetically modified *Sinorhizobium meliloti*. Soil Biology and Biochemistry 35 667–675.
- Ekman DR *et al.* (2003) SAGE analysis of transcriptome responses in Arabidopsis roots exposed to 2,4,6–trinitrotoluene. Plant Physiology 133 1397–1406.
- French CE *et al.* (1999) Biodegradation of explosives by transgenic plants expressing pentaerythritol tetranitrate reductase. Nature Biotechnology 17 491–494.
- Garg R *et al.* (1991) Treatment of explosives contaminated lagoon sludge. Hazardous Waste and Hazardous Materials 8 319–340.
- Gogal RM *et al.* (2002) Influence of dietary 2,4,6-trinitrotoluene exposure in the northern bobwhite (*Colinus virginianus*). Environmental Toxicology and Chemistry 21 81–86.
- Gong P *et al.* (2002) Toxicity of octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine (HMX) to soil microbes. Bulletin of Environmental Contamination and Toxicology 69 97–103.
- Gong P *et al.* (1999) Soil-based phytotoxicity of 2,4,6-trinitrotoluene (TNT) to terrestrial higher plants. Archives of Environmental Contamination and Toxicology 36 152–157.
- Hannink NK *et al.* (2001) Phytodetoxification of TNT by transgenic plants expressing a bacterial nitroreductase. Nature Biotechnology 19 1168–1172
- Hannink NK *et al.* (2002) Phytoremediation of explosives. Critical Reviews in Plant Sciences 21 511–538.
- Harvey SD *et al.* (1991) Fate of the explosive hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX) in soil and bioaccumulation in bush bean hydroponic plants. Environmental Toxicology and Chemistry 10 845–855.
- HSDB (2000) Hazardous Substances Data Bank (http://toxnet.nlm.nih.gov).
- Hughes JB *et al.* (1999) Anaerobic biotransformation of 2,4-dinitrotoluene and 2,6-dinitrotoluene by Clostridium acetobutylicum: A pathway through dihydroxylamino intermediates. Environmental Science and Technology 33 1065–1070.
- Just CL Schnoor JL (2000) A preparation technique for analysis of explosives in plant tissues. International Journal of Phytoremediation 2 255–267.
- Just CL Schnoor JL (2004) Phytophotolysis of hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX) in leaves of reed canary grass. Environmental Science and Technology 38 290–295.
- Komoba D *et al.* (1995) Metabolic processes for organic chemicals in plants. In: Plant Contamination: Modeling and Simulation of Organic Chemical Processes (Trapp S McFarlane JC eds.). Boca Raton: CRC Press Inc., pp 69–103.
- Lotufo GR et al. (2001) Toxicity of sediment-associated nitroaromatic and cyclonitramine compounds to benthic invertebrates. Environmental Toxicology and Chemistry 20 1762–1771.
- Medina VF *et al.* (2003) Plant tolerances to contaminants. In: Phytoremediation: Transformation and control of contaminants (McCutcheon SC Schnoor JL eds.). New Jersey: Wiley-Interscience Inc., pp 189–232.

- Mentewab A *et al.* (2005) Genomic analysis of the response of *Arabidopsis thaliana* to trinitrotoluene as revealed by cDNA microarrays. Plant Science 168 1409–1424.
- Mitchell WR Burrows EP (1995) Nitroreduction of 2,4-dinitrotoluene in-vitro by cytochrome P450 induced H4iie cells. Chemosphere 31 2767–2777.
- Miya RK Firestone MK (2001) Enhanced phenanthrene biodegradation in soil by slender oat root exudates and root debris. Journal of Environmental Quality 30 1911–1918.
- Moon H et al. (2004) Studies of TNT toxicity to plants using *Arabidopsis* as a model system. Plant Biology Annual Meeting Abstract 55. http://abstracts.aspb.org/pb2004/public/P31/7691.html.
- Newman LA *et al.* (1997) Uptake and biotransformation of trichloroethylene by hybrid poplars. Environmental Science and Technology 31 1062–1067.
- Nishino SF *et al.* (1999) Mineralization of 2,4- and 2,6-dinitrotoluene in soil slurries. Environmental Science and Technology 33 1060–1064.
- Nishino SF *et al.* (2000) Aerobic degradation of dinitrotoluenes and pathway for bacterial degradation of 2,6–dinitrotoluene. Applied and Environmental Microbiology 66 2139–2147.
- Padda RS *et al.* (2003) Mutagenicity of nitroaromatic degradation compounds. Environmental Toxicology and Chemistry 22 2293–2297.
- Palazzo AJ Leggett DC (1986) Effect and disposition of TNT in a terrestrial plant. Journal of Environmental Quality 15 49–52.
- Paquin D *et al.* (2002) Bench-scale phytoremediation of polycyclic aromatic hydrocarbon–contaminated marine sediment with tropical plants. International Journal of Phytoremediation 4 297–313.
- Pavlostathis SG *et al.* (1998) Transformation of 2,4,6-trinitrotoluene by the aquatic plant Myriophyllum spicatum. Environmental Toxicology and Chemistry 17 2266–2273.
- Picka K Friedl Z (2004) Phytotoxicity of some toluene nitroderivatives and products of their reduction. Fresenius Environmental Bulletin 13 789–794.
- Robidoux PY et al. (2000) Chronic toxicity of energetic compounds in soil determined using the earthworm (*Eisenia andrei*) reproduction test. Environmental Toxicology and Chemistry 19 1764–1773.
- Robidoux PY *et al.* (2001) Chronic toxicity of octahydro-1,3,5,7-tetranito-1,3,57-tetrazocine (HMX) in soil determined using the earthworm (*Eisenia andrei*) reproduction test. Environmental Pollution 111 283–292.
- Rodgers JD Bunce NJ (2001) Treatment methods for the remediation of nitroaromatic explosives. Water Research 35 2101–2111.
- Rosser SJ *et al.* (2001) Engineering plants for the phytodetoxification of explosives. In Vitro Cellular and Developmental Biology–Plant 37 330–333.
- Salt DE *et al.* (1998) Phytoremediation. Annual Review of Plant Physiology and Plant Molecular Biology 49 643–668.
- Sandermann H (1994) Higher plant metabolism of xenobiotics: the green liver concept. Pharmacogenetics 4 225–241.
- Schnoor JL *et al.* (1995) Phytoremediation of organic and nutrient contaminants. Environmental Science Technology 29 A318–A323.

- Schroder P Collins C (2002) Conjugating enzymes involved in xenobiotic metabolism of organic xenobiotics in plants. International Journal of Phytoremediation 4 247–265.
- Sens C *et al.* (1998) Distribution of C-14-TNT and derivatives in different biochemical compartments of *Phaseolus vulgaris*. Environmental Science and Pollution Research 5 202–208.
- Steevens JA *et al.* (2002) Toxicity of the explosives 2,4,6-trinitrotoluene, hexahydro-1,3,5-trinitro-1,3,5-triazine, and octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine in sediments to Chironomus tentans and Hyalella azteca: Low-dose hormesis and high-dose mortality. Environmental Toxicology and Chemistry 21 1475–1482.
- Subramanian M (2004) TNT Phytotransformation in Native and Genetically Modified Species: Significance of Aromatic Hydroxylamines in the Metabolic Pathway. PhD Thesis: Iowa State University.
- Subramanian M Shanks JV (2003) Role of plants in the transformation of explosives. In: Phytoremediation: Transformation and control of contaminants. (McCutcheon SC Schnoor JL eds.). New Jersey: Wiley-Interscience Inc., pp 389–408.
- Subramanian M *et al.* (2005) TNT phytotransformation pathway characteristics in Arabidopsis: Role of aromatic hydroxylamines. Biotechnology Progress (in press).
- Talmage SS *et al.* (1999) Nitroaromatic munition compounds:Environmental effects and screening values. Reviews of Environmental Contamination and Toxicology 161 1–156.
- Thompson PL (1997) Phytoremedation of Munitions (RDX, TNT) Waste at the Iowa Army Ammunition Plant with Hybrid Poplar Trees. PhD Thesis: The University of Iowa.
- Thompson PL *et al* (1998) Uptake and transformation of TNT by hybrid poplar trees. Environmental Science and Technology 32 975–980.
- Thompson PL *et al* (1999) Hexahydro-1,3,5-trinitro-1,3,5-triazine translocation in poplar trees. Environmental Toxicology and Chemistry 18 279–284.
- Todd SR Lange CR (1996) Phytoremediation of 2,4-dinitrotoluene contaminated soils using parrot feather (*Myriophilum brasiliense*). Hazardous and Industrial Wastes Proceedings of the Mid-Atlantic Industrial Waste Conference 557–564.
- Van Aken B *et al.* (2004) Metabolism and mineralization of hexahydro-1,3,5-trinitro-1,3,5-triazine inside poplar tissues (*Populus deltoides* × *nigra* DN-34). Environmental Science and Technology 38 4572–4579.
- Vervaeke P *et al.* (2003) Phytoremediation prospects of willow stands on contaminated sediment: a field trial. Environmental Pollution 126 275–282.
- Vila M *et al.* (2005) Metabolism of C-14-2,4,6-trinitrotoluene in tobacco cell suspension cultures. Environmental Science and Technology 39 663–672.
- Wang C Hughes JB (1998) Derivatization and separation of 2,4,6–trinitrotoluene metabolic products. Biotechnology Techniques 12 839–842.
- Wang CY *et al.* (2003) Role of hydroxylamine intermediates in the phytotransformation of 2,4,6-trinitrotoluene by *Myriophyllum aquaticum*. Environmental Science and Technology 37 3595–3600.
- Yinon J Zitrin S (1993) Modern Methods and Applications in Analysis of Explosives. New York: John Wiley and Sons.



JACQUELINE SHANKS is a professor of chemical engineering at Iowa State University (ISU) and an adjunct professor of bioengineering at Rice University. She received her BS from ISU in 1983 and her PhD from the California Institute of Technology in 1989. She joined Rice in 1988 and ISU 1999.

Dr. Shanks' research interests include engineering of secondary metabolites in plants, nuclear magnetic resonance spectroscopy techniques for metabolic flux analysis, phytoremediation of explosives and related

nitroaromatics, and production of valuable products from biorenewable resources.

She received the NSF Young Investigator Award in 1992 and ISU's Professional Progress in Engineering Award in 1994. She was elected as fellow to the American Institute of Medical and Biological Engineers in 2000, and served as a member of the NRC Committee on Biobased Industrial Products. She is a member of the ACS's Biochemical Technology (BIOT) and Environmental Chemistry (ENVI) Divisions. She received the Van Lanen Award for service in the BIOT division in 2004. Shanks has served as co-editor of the Biochemical Engineering section of Current Opinion in Biotechnology and as co-editor for a 2002 issue of Metabolic Engineering devoted to plant metabolic engineering. She is a member of the Editorial Advisory Board for Biotechnology Progress.